

This is a post-peer-review, pre-copyedit version of an article published in Forest Ecology and Management.

The final authenticated version is available online at:  
<https://doi.org/10.1016/j.foreco.2019.06.031>

Terms of use: see Elsevier <https://www.elsevier.com/about/policies/sharing> for archived author accepted manuscripts (AAMs) of subscription articles.

Datoteka je postprint, avtorjeva končna, recenzirana različica članka, sprejeta v objavo v reviji Forest Ecology and Management.

Končna, originalna založnikova verzija je na voljo na spletu:  
<https://doi.org/10.1016/j.foreco.2019.06.031>

Pogoji uporabe: glej Elsevier <https://www.elsevier.com/about/policies/sharing> pogoje ponovne uporabe za arhivirane, avtorjeve končne, recenzirane različice članka (AAMs), objavljene v naročniški reviji.

1 Growing stock of nectar- and honeydew-producing tree species

2 determines the beekeepers' profit

3

4 Authors: Janez Prešern, Agricultural Institute of Slovenia

5 Jan Mihelič, University of Ljubljana, Dept. of Forestry

6 Milan Kobal, University of Ljubljana, Dept. of Forestry

7

8

## 9 Abstract

10

11 Forests and woodlands are considered as the most important sources of honey bee  
12 forage in many European countries with several tree species providing nectar and/or  
13 honeydew flow. Slovenia boasts with high number of beekeepers and high colony count  
14 per square kilometer. We have investigated the impact of availability of natural  
15 resources and colony density on honey yield.

16

17 Data presented here were collected on 57 locations with monitor hives, equipped with  
18 scales, over years 2011 – 2016. Locations were selected according to site vegetation,  
19 ensuring identified source of nectar or honeydew flow. The source of the flow was  
20 recorded and verified by contract beekeeper. We investigated 1) the relationship  
21 between abundance of the flow source expressed as the quantity of growing stock and  
22 net mass gain of the monitor colony during the flow and 2) the relationship between  
23 colony density expressed as the number of colonies against growing stock volume and  
24 net mass gain of the monitor colonies.

25

26 We found an asymptotic exponential relationship between colony mass gain and  
27 growing stock of the species, providing flow. The exception was the spruce where the  
28 relationship was determined as linear ( $k = 0.023 \pm 0.009$ ). The  $\tau$  of the exponential  
29 approach in the case of acacia flow was  $9.8 \pm 5.6$  and in case of linden flow  $6.6 \pm 3.9$   
30 (mean  $\pm$  SE). Colony density then determined the colony mass gain due to flow. In  
31 cases of acacia, linden and spruce flow we have determined the relationship between  
32 colony density and mass gain as decaying exponential ( $\tau = 283.9 \pm 60.6$ ,  $\tau = 1.6 \pm 0.4$   
33 and  $3.0 \pm 1.3$ , respectively, all mean  $\pm$  SE). Combined linden/chestnut flow was fitted

34 best with linear equation ( $k = - 0.08 \pm 0.019$ ). Most likely, another variable should be  
35 used in the case of spruce flow: population of dew-producing insects. Periodical  
36 monitoring of eight acacia locations show differences in mass gain between years, thus  
37 allowing prediction of colony densities which guarantee profit: these locations are  
38 determined as those with colony density less than 50 hives/10<sup>3</sup> m<sup>3</sup> growing stock gained  
39 more than 10 kg/hive in 83% of cases, regardless of the year.

40

41 Our results indicate that a cap on the total number of colonies at one location should be  
42 considered to maximize beekeepers' profit.

43

44

45 **Keywords:** beekeeping, forage, environment carrying capacity, non-wood forest  
46 products

## 47 Introduction

48

49 Forests and woodlands are considered as the most important sources of honey in many  
50 European countries with several tree species providing nectar and/or honeydew flow.  
51 Even in regions like sub-Saharan Africa, honey has already been recognized as valuable  
52 non-wood forest product (Chikamai et al. 2009). Beekeepers often pursue single-source  
53 flows – e.g. linden bloom or spruce honeydew - to obtain honey with distinct sensorical  
54 properties (c. f. Persano Oddo & Piro 2004, Crane and Walker 1985) - such honeys  
55 achieve higher prices on the market. Beside linden and spruce, honeys like acacia,  
56 chestnut and fir honey are recognized as important single-source honeys in Slovenia  
57 and neighboring countries. In contrast with many EU and non-EU countries, cultured  
58 plants are of lesser importance for honey production in Slovenia (for recognized honey  
59 sorts see Official Gazette of the Republic of Slovenia, 2009). Desire to maximize honey  
60 yield calls for potential regulation of the colony density in forest stands in which the  
61 nectar/dew flow is expected. Such a move may be strongly opposed by beekeeping  
62 community, yet data is very scarce. To uphold the regulation a thorough analysis should  
63 be made, and a model developed.

64

65 Honey bees are forest dwellers, establishing colonies in hollow tree trunks. Natural  
66 colony density as determined for both Palearctic and Nearctic forests was established  
67 at 0.5 colonies/km<sup>2</sup> (Galton, 1971, Visscher & Seeley, 1982). With the arrival of deadly  
68 ectoparasite *Varroa destructor*, feral colonies were thought to be mostly extinct.  
69 However, Kohl & Rutschmann (2018) showed a contemporary natural density of 0.11–  
70 0.14 honey bee colonies/km<sup>2</sup> in two European beech forests of Germany, which is in  
71 agreement with 0.1 colonies/km<sup>2</sup> reported from Poland (Oleksa, Gawroński & Tofilski,

72 2013) and an order of magnitude less than 1.0 colonies/km<sup>2</sup> in later census in a mature  
73 Nearctic hardwood forest (Seeley, 2007). These densities are fairly low compared to  
74 the densities of managed colonies reported by national registries for certain countries.  
75 For example, there were roughly 160.000 registered colonies on Slovenia's territory of  
76 20.273 km<sup>2</sup> in the years up till 2017, giving densities around 8 colonies/km<sup>2</sup>. Most of  
77 the colonies (>99%) are registered under 900 m above sea level, taking higher  
78 mountainous regions (15 % of total surface) out of the equation. The recalculated  
79 density is then 8.64 colonies/km<sup>2</sup> on 11.764 locations, giving an average a bit more than  
80 13 colonies/apiary. Official honey production varies from year to year with maximum  
81 of 2047 tons or 12.8 kg/colony in 2015 (Statistical Office RS, 2018).

82

83 The natural dispersal of a colony's foragers over the space has been often addressed. In  
84 cases with low colony density, a skewed distribution of foragers was reported with  
85 median forage range of 1650 m and maximum range of 10,100 m (Visscher & Seeley  
86 1982). A similar report was given by Steffan-Dewenter and Kuhn (2003) who noted  
87 median range of 1,181 m and maximum range of 10,037 m. In cases where food patches  
88 are more abundant, Waddington et al. (1994) showed that distances and forage pattern  
89 can differ significantly between locations and colonies, yet most of foraging happens  
90 within 3 km radius. However, in situations with high colony density foragers disperse  
91 to places burdened with lesser colony numbers (Gary 1978).

92

93 The Slovenian government-sanctioned Queen breeding program for *Apis mellifera*  
94 *carnica* contains provisions for monitoring of the natural resources, namely nectar and  
95 honeydew flow monitoring. Monitoring service was established more than 30 years ago  
96 and is currently being managed by experts at Slovenian Beekeepers' Association

97 (SBA). It consists of a network of monitoring colonies (50 – 70, depending on the year),  
98 equipped with manual and automatic hive scales at selected locations, which provide  
99 certain kind of nectar or honeydew flow (e.g. acacia, linden, spruce, fir...). Contract  
100 beekeepers maintain the monitoring colonies and confirm the type of the flow.  
101 Selection of locations is based on the geodatabase of forest stands (Slovenia Forest  
102 Service, 2015) and monitoring growing stock for different tree species every 10 years.

103

104 In this paper we discuss growing stock as determinant of the resource availability and  
105 the colony density's bearing on colony mass gain and colony honey production.



## 106 Methods

107

### 108 Data sources

109

110 Presented data about nectar and honeydew flow were collected in years 2011 – 2016  
111 with monitoring hives in care of contract beekeepers. Mass changes of these hives were  
112 recorded daily by means of commercially-available automatic GSM-equipped hive  
113 scales (Ames d.o.o and Eldema d.o.o, both Slovenia). The data collection was and still  
114 is organized and coordinated organized by Slovenian Beekeepers' Association; daily  
115 changes are available to public via online portal (<https://ecelebar.czs.si>). All  
116 measurements collected are stored in the database maintained at the Agricultural  
117 Institute of Slovenia. We have extracted daily data for selected acacia, linden, mixed  
118 linden/chestnut nectar flows and spruce honeydew flow. While botanical sources of  
119 linden, chestnut and spruce flows are clear (*Tilia spp.*, *Castanea sativa*, *Picea abies*,  
120 correspondingly), the acacia nectar flow is actually provided by black locust tree  
121 (*Robinia pseudacacia*) which in beekeeping community is termed “acacia”. We will  
122 maintain this term through the paper when discussing flow and/or honey. Type of the  
123 flow is verified by contract beekeeper who took care of the hive. Determination of the  
124 source was based on timing, vegetation around the hive and sensory qualities. No  
125 palynological validation was performed. In cases where overlap in time was possible  
126 (linden/chestnut), care was taken either to exclude any doubt about the flow purity or  
127 to assign the measurements to mixed category. Monitoring hives are strategically  
128 placed at the locations interesting for beekeepers (expert judgement, Fig 1A).  
129 Depending on the situation, some locations are kept for years, others are changed yearly  
130 (Table 1). Beekeepers must report establishment of a new apiary as well colony counts

131 at the apiaries under their care; colony counts are collected twice a year. Apiary  
132 coordinates and corresponding yearly colony counts were obtained from database of  
133 national veterinary administration (UVHVVR). We have counted colonies within 3 km  
134 radius of the monitoring colonies, using custom-written Python script.

135

136 Data about growing stock of certain tree species within a forest stand around apiary (in  
137 m<sup>3</sup>) were calculated based on the data of forest stand map, produced every ten years by  
138 Slovenian Forest Service for forest management plans. Based on remote sensing data  
139 (orthophoto images at 1:5.000 scale) and field survey, homogenous forest stands were  
140 delineated (Slovenian Forest Service, 2015). In each delineated forest stand, several  
141 temporary sampling plots (minimum 7 per stand) are established during field work. The  
142 growing stock for each tree species in each plot was estimated with Bitterlich's angle-  
143 count method and measurement of average tree height per different tree species  
144 (Bitterlich 1948). Finally, sample plot data were averaged for each forest stand and  
145 growing stock per tree species (in m<sup>3</sup>/ha) was calculated. To obtain information on  
146 growing stock per tree species per stand level (in m<sup>3</sup>), the growing stock of certain tree  
147 species within a forest stand was calculated according to the area of specific forest  
148 stand.

149

150 Growing stocks per tree species were later evaluated in 3 km radius around the  
151 monitoring colonies. Both, forest stand map and apiary coordinates were imported into  
152 GIS (ArcMAP 10.6; ESRI, 2018) and only forest stands within 3 km buffer (function  
153 *Buffer*) around each apiary was selected for further analyses (function *Clip*). For forest  
154 stands on border of 3 km, growing stock per tree species per stand level were corrected  
155 according to the proportion of the forest stand inside / outside buffer of 3 km. Growing

156 stock per tree species inside buffer of 3 km was finally calculated by summarizing data  
157 from selected forest stands (Fig 1B).

158

159 We have defined colony density as a number of colonies per volume of growing stock  
160 of selected tree species to offset different amount of resources at different locations.

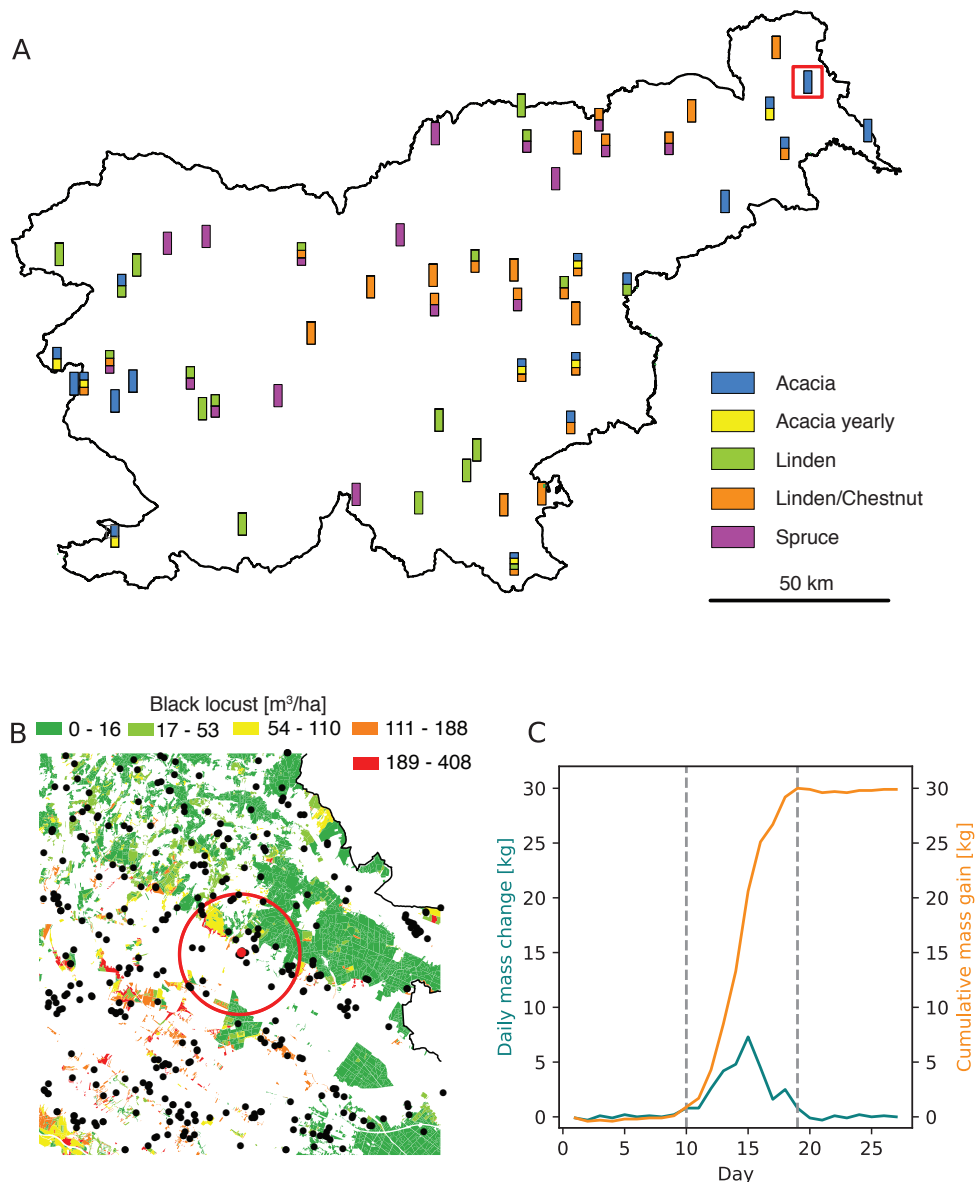
161 Colony density was calculated using custom-written Python script.

162

163 *Table 1. Number of monitor colonies per year for each type of flow. Time interval in which flow appears is also*  
164 *marked.*

	<b>Acacia</b>	<b>Linden</b>	<b>Linden/Chestnut</b>	<b>Spruce</b>	<b>N</b>
<b>2011</b>	13	12	9	8	42
<b>2012</b>	15	10	9	15	49
<b>2013</b>	14	9	12	5	40
<b>2014</b>	14	9	20	4	47
<b>2015</b>	15	10	19	3	47
<b>2016</b>	14	12	19	4	49
<b>N data</b>	<b>85</b>	<b>62</b>	<b>88</b>	<b>39</b>	<b>274</b>
<b>N unique locations</b>	<b>18</b>	<b>19</b>	<b>26</b>	<b>18</b>	
<b>Observed flow period</b>	end of April- end of May	end of May- early July	end of July	May-early July	end of April- early July

165



166

167 *Figure 1. A: locations of monitoring hives within Slovenia, providing data for this study. In several locations,*  
 168 *monitoring hive provided data for different nectar flows (see legend). Locations marked as “Acacia yearly” were*  
 169 *used in the analysis of yearly differences in acacia nectar flow. Location “Bogojina” marked with a red square. B:*  
 170 *Neighborhood of monitoring hive at location “Bogojina”. Red circle marks 3 km radius around the hive. Black dots*  
 171 *represent apiaries in the surroundings. Black locusts’ growing stocks are color coded. See legend within figure. C:*  
 172 *Daily mass changes (green) and cumulative mass gain (orange) at the location “Bogojina” in 2015 due to the acacia*  
 173 *nectar flow.*

174 Data analysis

175

176 The analysis was made for the selected time interval for relevant locations. Time  
177 interval of nectar flow was determined by both the beekeeper charged with care of the  
178 station and deviation of daily mass gain from the baseline (Fig 1C) during the time of  
179 bloom. Data within this interval was then summed for each location and each year.

180

181 We have investigated relationship of growing stock on monitoring colony mass gain,  
182 and the relationship between the colony density and the colony mass gain. Python's  
183 *lmfit* package was used in our custom-written scripts (DOI: 10.5281/zenodo.3248183)  
184 to perform fits and evaluate the quality of the fit.

185

186 We describe the relationship between quantities by fitting the recorded values by either  
187 exponential approach (Eq 1; in case of growing stock, custom-written) or exponential  
188 decay for colony density (Eq 2; included in *lmfit* package) using least-squares  
189 minimization (LSM) method. Alternatively, a linear regression built in *lmfit* package  
190 was used, using LSM as well.

191

192 Eq 1: 
$$y = A \times (1 - e^{-\frac{x}{\tau}}) + B$$

193

194 Eq 2: 
$$y = A \times e^{-\tau \times x} + B$$

195

196 To decide between the exponential or linear approach we used Akaike information  
197 criterion (AIC), implemented in *lmfit* package, selecting the model with lower AIC

198 score. Goodness of fit is reported with standard error, also implemented within *lmfit*  
199 package.

## 200 Results

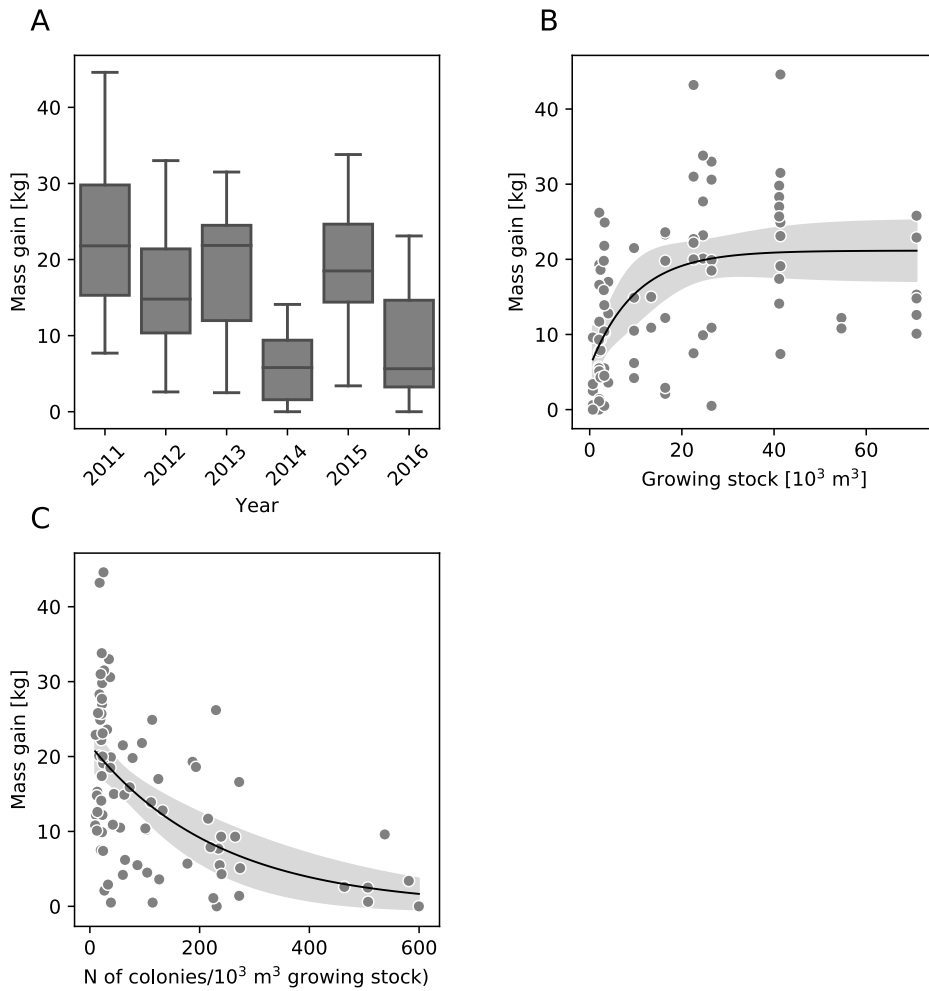
201

### 202 Acacia nectar flow

203

204 Flow begins in late April/early May - depending on the geographical location - and lasts  
205 between 2 and 16 days in years observed. Eighty-five measurements were collected at  
206 18 locations over the years 2011 – 2016 (Table 1). Range of mass gain during acacia  
207 nectar flow in the years 2011 – 2016 was between 0.0 and 44.6 kg, with median 14.1  
208 kg and interquartile range of 15.3 kg. First and most evident thing are the differences  
209 in median mass gain between years: from 5.8 kg in 2014 to 21.85 kg in 2013. We  
210 present values in Fig 2A. Average yearly durations of the flow were between 5.2 and  
211 6.9 days with the exception of 2014, in which the flow was cut short to 2.4 days.

212



213

214 *Figure 2. A: Mass gains during acacia nectar flow. Boxplots show huge variability between years. B: Impact of*  
 215 *black locust's growing stocks on measured mass gain. Figure shows mass gains reaching ceiling at approximately*  
 216  *$40 \times 10^3 \text{ m}^3$  of growing stocks ( $\tau = 9.8 \pm 5.6$ , mean  $\pm$  SE). C: Mass gain during nectar flow depends on colony*  
 217 *density. Colony density was computed as number of hives on  $10^3 \times \text{m}^3$  of black locust growing stocks ( $\tau = 283.9 \pm$*   
 218 *60.6). Each dot in B and C represents a single data point. Shaded areas in B and C mark 95% confidence interval.*

219 Relationship between *Robinia's* growing stock and colony mass gain during acacia  
 220 nectar flow was best described by exponential approach function (Eq 1;  $\tau = 9.8 \pm 5.9$ ,  
 221 mean  $\pm$  SE;  $\text{AIC}_{\text{expapp}} = 370$  vs.  $\text{AIC}_{\text{lin}} = 383$ ), reaching ceiling around  $200.000 \text{ m}^3$  of  
 222 growing stock (Fig 2B). Finally, Fig 2C shows importance of colony density. Locations  
 223 burdened with more than 200 colonies per  $1000 \text{ m}^3$  of black locust growing stock

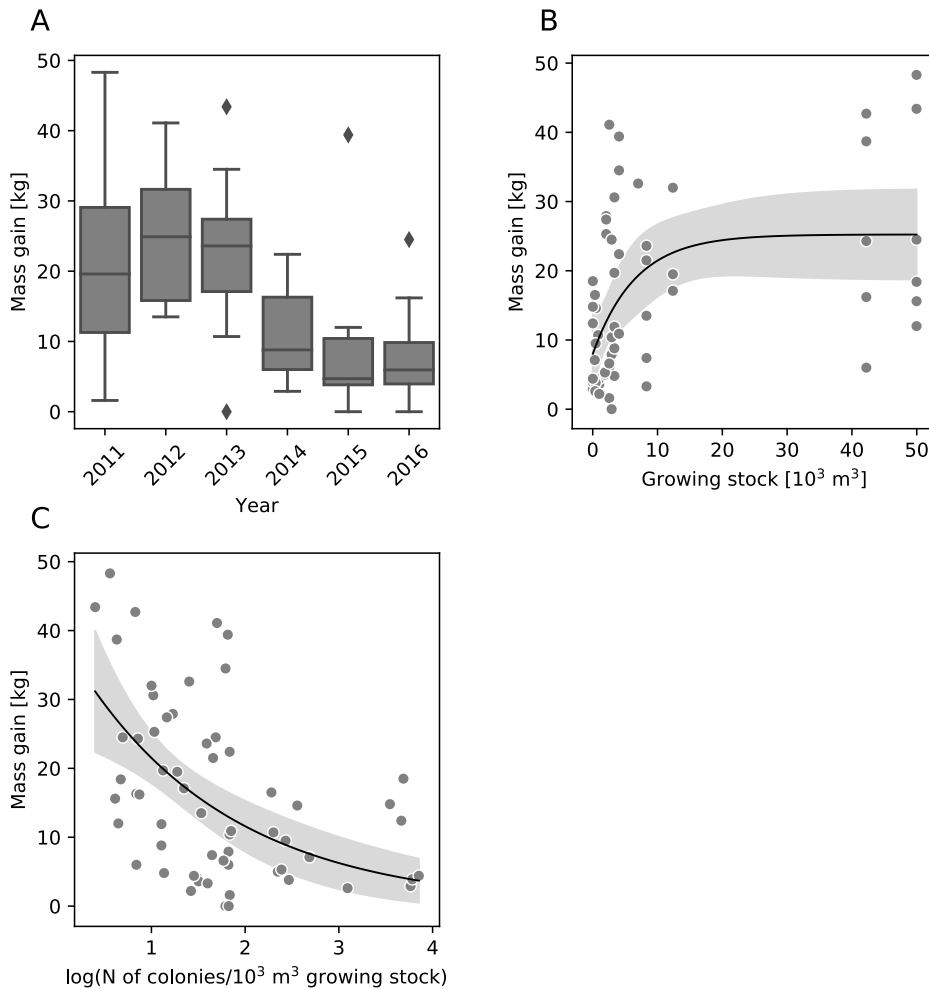


224 gained more than 10 kg only in 3 cases (16%) and locations, burdened with 400 colonies  
225 per 1000 m<sup>3</sup> without single exception harvested less than 10 kg/hive. The relationship  
226 between density and colony mass gain is not linear either: exponential decay fits best  
227 (Eq 2;  $\tau = 233.9 \pm 60.6$ , mean  $\pm$  SE;  $AIC_{\text{expdec}} = 368$  vs.  $AIC_{\text{lin}} = 372$ ). Locations  
228 burdened with less than 50 hives/1000 m<sup>3</sup> had harvested 10 kg/hive and more in 85 %  
229 of cases.

230

231 Linden and mixed linden/chestnut nectar flow

232



233

234 *Figure 3. A: Colony mass gain during linden nectar flow. Boxplots show variability between years. B: Dependence*  
 235 *of mass gain on linden growing stock ( $\tau = 6.6 \pm 3.9$ , mean  $\pm$  SE). C: Dependence of mass gain on log colony density*  
 236 *( $\tau = 1.6 \pm 0.4$ , mean  $\pm$  SE). Each dot in B and C represents a single data point. Shaded areas in B and C mark 95%*  
 237 *confidence interval.*

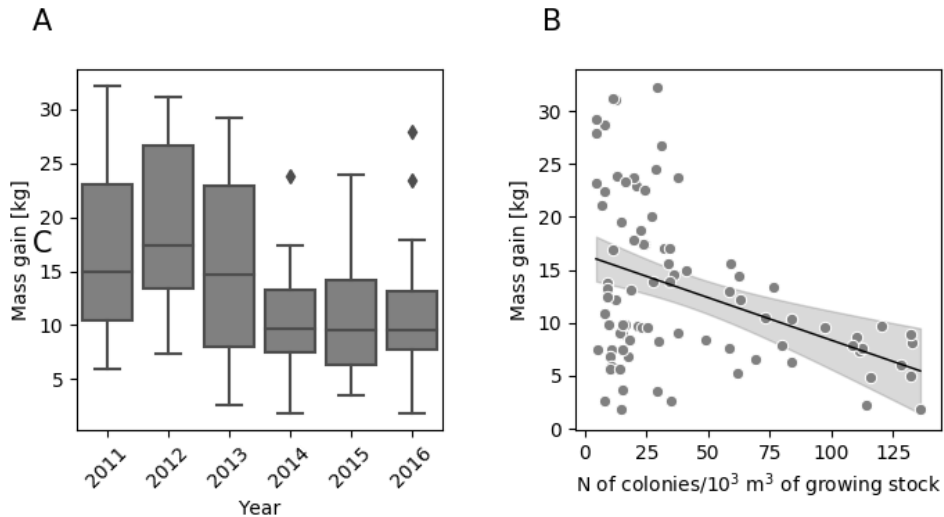
238 Linden nectar flow (*Tilia* spp.) is usually due in June and in many locations overlaps at  
 239 least partially with chestnut (*Castanea sativa*) flow. We took data from linden-only  
 240 locations and combined locations for years in which flow did not overlap, all together  
 241 62 measurements from 19 locations (Table 1). Nectar flow differed between years, with  
 242 overall median being 12.95 kg and interquartile range 19.05 kg (Fig 3A). Asymptotic  
 243 exponential approach offered best description between mass gain and growing stock

244 (Eq 1;  $\tau = 6.6 \pm 3.9$ , mean  $\pm$  SE;  $AIC_{\text{expapp}} = 303$  vs.  $AIC_{\text{lin}} = 305$ ; Fig 3B). To examine  
245 the impact of hive density on mass gain during linden flow, we have log-transformed  
246 colony density and fitted data with exponential decay (Eq 2;  $\tau = 1.6 \pm 0.4$ , mean  $\pm$  SE;  
247  $AIC_{\text{expdec}} = 299$  vs.  $AIC_{\text{lin}} = 304$ ; Fig 3C).

248

249 Overlapping of linden and chestnut nectar flow was recorded in 88 occasions in 26  
250 locations during last six years (Table 1). Again, mass gains of the monitoring colonies  
251 differed between years: median mass gain of years 2014, 2015 and 2016 combined was  
252 only 59 % (9.65 kg) compared to combined median mass gain recorded in years 2011,  
253 2012, 2013 (16.35 kg; Fig 4A). Colony density had negative impact on the mass gain  
254 of the monitoring colonies: negative trend due to higher colony density is evident in  
255 figure Fig 4B. Linear regression fitted data points ( $R^2 = 0.17$ ;  $k = -0.08 \pm 0.019$ , mean  
256  $\pm$  SE;). Both the exponential decay and linear model had similar AIC values (345 and  
257 345, respectively); we decided for more parsimonious linear model. Data recorded in  
258 locations with colony density higher than 50 colonies per 1000 m<sup>3</sup> gained more than 10  
259 kg of mass only in 8 out of 26 cases. With colony densities less than 50 per 1000 m<sup>3</sup> of  
260 wood, 37 out of 62 gained more than 10 kg mass during the flow.

261



262

263 *Figure 4. A: Mass gains during combined linden/chestnut nectar flow. Boxplots show large variability between*

264 *years. B: Increase of colony density shows decrease of colony mass gain ( $R^2 = 0.17$ ;  $k = -0.08 \pm 0.019$ , mean  $\pm$  SE).*

265 *Each dot in B represents a single data point. Shaded area in B marks 95% confidence interval.*

## 266 Spruce honeydew

267

268 Spruce honey is one of the two types of single-source honeydew honeys recognized in

269 Slovenia. 39 data points collected in years from 2011 to 2016 at 18 locations are shown

270 (Table 1). Median mass gains of the monitor colonies range from 11.4 kg in 2015 and

271 20.9 kg in 2010 (Fig 5A).

272

273 Spruce is one of the most important tree species in Slovenian forests, yet the

274 relationship between spruce growing stock and mass gain of monitor colonies is

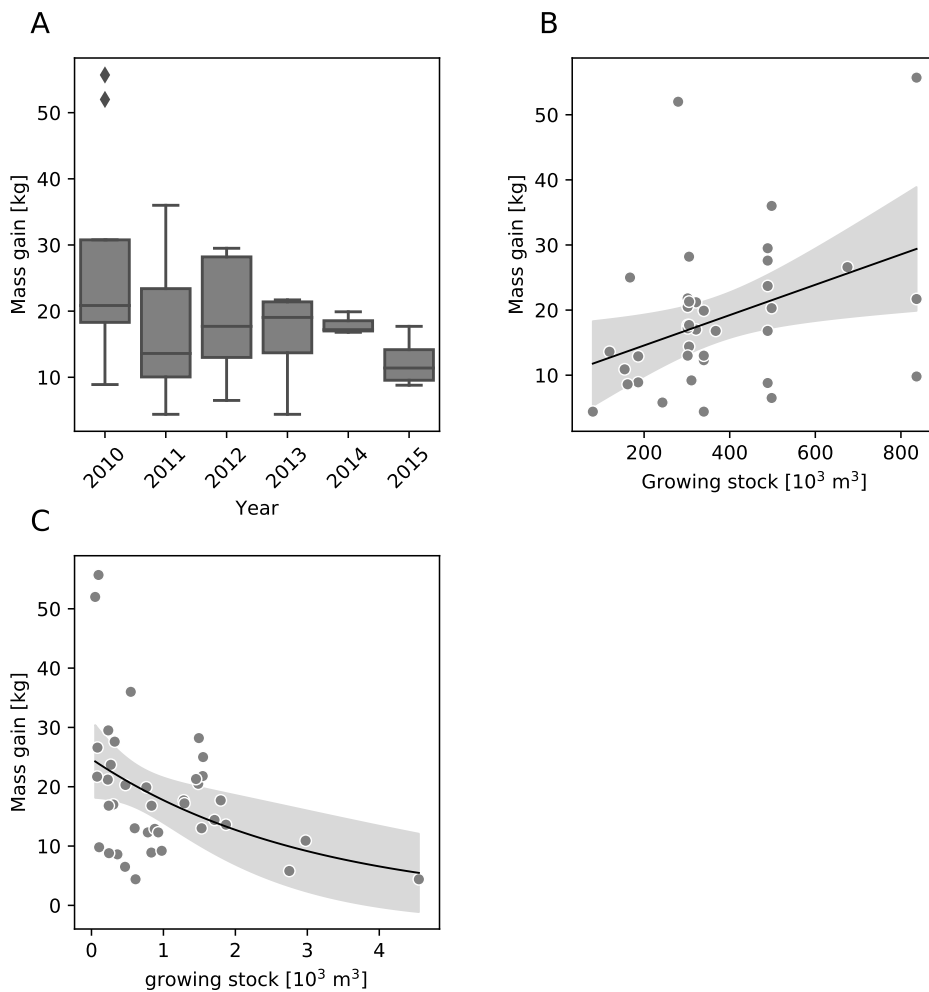
275 moderate at best. AIC values suggested use of linear model ( $AIC_{\text{expdec}} = 187$  vs.  $AIC_{\text{lin}}$

276  $= 185$ ); we described relationship by a linear equation ( $R^2 = 0.15$ ;  $k = 0.023 \pm 0.009$ ,

277 mean  $\pm$  SE; Fig 5B). Relationship between colony density and monitor colonies' mass

278 gain was described by decay exponential (Eq 2;  $\tau = 3.02 \pm 1.4$ , mean  $\pm$  SE;  $AIC_{\text{expapp}} =$

279  $184$  vs.  $AIC_{\text{lin}} = 185$ ; Fig 5C).



281

282 *Figure 5: Colony mass gain during spruce honeydew flow. Boxplots show variability between years. B: Dependence*  
 283 *of mass gain on spruce growing stock ( $R^2 = 0.15$ ;  $k = 0.023 \pm 0.009$ , mean  $\pm$  SE). C: Dependence of mass gain*  
 284 *colony density ( $\tau = 3.0 \pm 1.3$ , mean  $\pm$  SE). Each dot in B and C represents a single data point. Shaded areas in B*  
 285 *and C mark 95% confidence interval.*

## 286 Yearly differences in acacia nectar flow

287

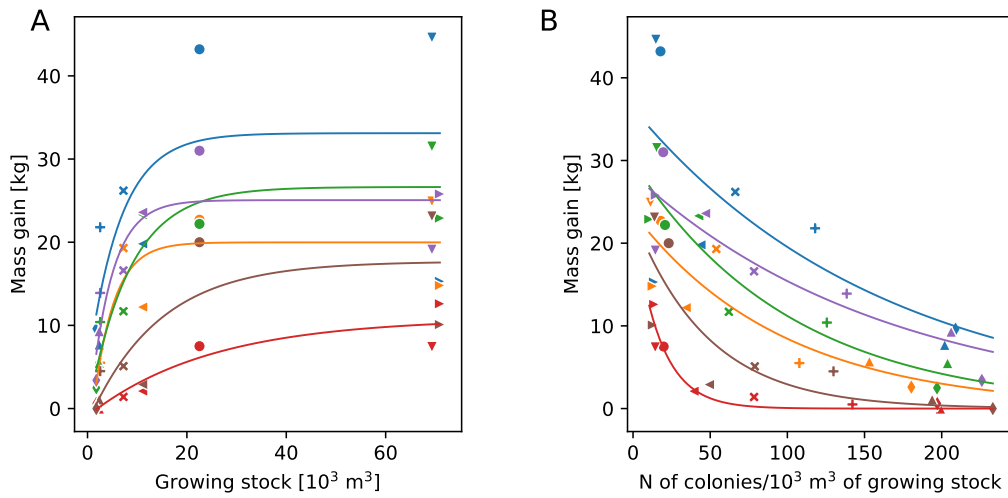
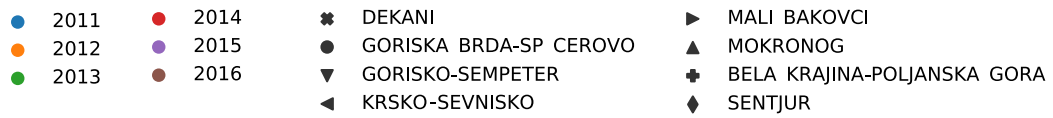
288 Eight out of eighteen acacia locations were monitored in all six years. We used this  
 289 subset for detailed analysis of inter-seasonal variation. In years with normal or good  
 290 nectar flow (2011, 2012, 2013 and 2015), an approach to ceiling could be observed

291 regardless of the growing stock. Exponential approach (Eq 1) reached ceiling regardless  
292 of black locust density and limiting honey yield at around 200,000 m<sup>3</sup> of black locust  
293 stocks ( $\tau = 4.1 \pm 3.3$  in 2012, mean  $\pm$  SE ). In years with low nectar flow however, the  
294 relationship is still asymptotic and relatively flat ( $\tau = 24.6 \pm 15.4$  in 2014, mean  $\pm$  SE),  
295 almost linear (Fig 6A).

296

297 Colonies in the environment burdened with less than 50 colonies per 1000 m<sup>3</sup> of black  
298 locust growing stock gained more than 10 kg/hive in 83 % of cases. On the other hand,  
299 colonies in the locations burdened with more than 200 colonies per 1000 m<sup>3</sup> of black  
300 locust growing stock would not gain more than 10 kg mass during the flow, regardless  
301 of the year. The maximum mass gain range was 37.2 kg in one of the locations  
302 (“Šempeter”), while minimum mass gain range was 9.3 kg in location “Mokronog”.  
303 The relationship between density and colony mass gain was described by exponential  
304 decay:  $\tau$  values were between  $17.2 \pm 6.5$  (2014) and  $164.1 \pm 43.9$  (2015; mean  $\pm$  SE;  
305 Eq 2).

306



307

308 *Figure 6. Mass gain due to acacia nectar flow, recorded every year on eight locations. A: Dependence of mass gain*

309 *due growing stock in 3 km range. Mass gain reaches ceiling in “normal” years regardless of available growing*

310 *stock. B: Mass gain during nectar flow depends on colony density. Colony density was computed as number of hives*

311 *on  $10^3 \times \text{m}^3$  of R. pseudacacia growing stocks. Each dot in A and B represents a single data point.*

312

313

## 314 Discussion

315

316 Beekeeping is very much woven into the fabric of the Slovenian nation. Consequently,  
317 a lot has been invested in research of various aspects of nectar and honeydew flows in  
318 the past. Unfortunately, most of the collected data hasn't been digitized and/or  
319 centralized. We used the available data to show the importance of natural resource  
320 scarcity and the environment capacity for honey bee colony density in connection with  
321 availability of nectar and honeydew flow.

322

### 323 Colony mass changes

324

325 We have analyzed mass gains in the monitoring hive during the bloom, yet the observed  
326 gains were not due to the nectar flow only. Honey bees seem to condense collected  
327 nectar already on the return flight (Nicolson & Human, 2008), a change that does not  
328 register on hive scales. Commercial hive scale system usually reports mass once a day.  
329 Beside water evaporation, calculated mass changes include also consumption for  
330 colonies' own needs; these needs vary with the state of the colony (Brodschneider &  
331 Creilsheim, 2010). Non-linear relationship between stored honey and colony mass was  
332 determined as polynomial, variability between locations ascribed to differences in  
333 moisture and temperature (McLellan, 1977). It is possible to separate daily needs of the  
334 colony using smoothing average of hourly mass measurements as a reference line.  
335 Amplitudes of hourly changes against such background show colonies' physiological  
336 needs, water evaporation etc. It is possible to discern trends in gaining/losing food  
337 stores over longer period e.g. week (Meikle et al. 2008). For all these reasons it is



338 impossible to use linear relationship to predict honey harvest directly. Yet it is possible  
339 to use the measured mass changes to get rough ball-park estimation.

340

#### 341 Availability of resources

342

343 In this paper we use absolute growing stock (in m<sup>3</sup>) to describe available resources.  
344 Alternative to complex data of growing stock would be land use data e.g. percentage of  
345 forest area. While percentage of forest area might be directly comparable to growing  
346 stock in pure mono-cultures or forests with single forest tree species, percentage of  
347 forest area is much less reliable in diverse and natural forests of eastern and southern  
348 part of Europe (Pirnat, 2017).

349

350 Sites providing single-source flows are sought for – local single-source honeys have  
351 their unique characteristics and usually achieve higher market prices. Black locust (*R.*  
352 *pseudacacia*) blooms early in season, usually in late April or early May, giving so-  
353 called acacia honey. Honey potential for one hectare of black locust was reported to be  
354 between 48 and 1600 kg, range depending on the country (Crane et al., 1984). For  
355 consideration, 36 years old and very pure black locust forest stands can contain 282  
356 m<sup>3</sup>/ha (Redei and Meilby, 2000). A noteworthy observation (Fig 2B) is the existence  
357 of ceiling: increase of resource availability over 200.000 m<sup>3</sup> does not improve mass  
358 gain considerably. Linden trees (*Tilia* spp.) have similar honey potential as black locust,  
359 90 – 1000 kg, again, depending on the country (Jašmak, 1973; Crane et al., 1984);  
360 however, one should take into the account that linden trees are rarely dominant tree  
361 species in forest stands of Slovenia in contrast to black locust (Brus, 2012). The  
362 effective foraging radius in Europe was estimated to about 2 – 3 km (von Frisch 1965),

363 yet there is no way to tell how far from the hive did the colony forage. In the dense and  
364 floristically monotonous associations where black locust is dominant, and in which  
365 colony density is high, it is easy to assume dispersal of the foragers to places burdened  
366 with lesser colony numbers, like observed in the orchards (Gary 1978).

367

368 On the other hand, linden and chestnut trees are normally found in floristically diverse  
369 associations in which they do not represent major component. Due to sparsity of linden  
370 trees in such associations, one could assume that colonies forage in patch-like fashion,  
371 focusing on single patch and switching only when quality of the patch drops (c.f.  
372 Waddington et al., 1994). Chestnut grows on acidic soil as the main supporting species  
373 of several associations and does not form continuous forests (Brus, 2007).

374

375 Honeydews are dependent on insects sucking sap, adding another variable to the  
376 equation. Spruce trees host two important species *Physokermes piceae* and *P.*  
377 *hemicryphus*, which are one of the most important sources of dew. These insects are  
378 univoltine and overwinter as larvae, making them susceptible to weather conditions  
379 (Rihar, 1992). Proper explanation of mass gain during honeydew flow would then  
380 require knowledge of the insect's population and poor fits are likely consequence of the  
381 lack of knowledge about this component.

382

### 383 Colony density and economic consequences

384

385 Regardless of the flow source, our results show importance of colony density: higher  
386 the density, lower the mass gain. This is especially important in the seasons with weak  
387 nectar flow when the carrying capacity of the environment could be reached already at

388 low colony densities (Fig 6B, years 2014 and 2016). While it is difficult to predict a  
389 weak season without very evident reason like spring frost, it is clear that seasons differ  
390 in mass gains and honey yield. There seem to be several parameters, influencing nectar  
391 flow, among them weather conditions during the preceding winter. In case of acacia,  
392 the most important seem to be temperature requirements: > 25 °C during daytime and  
393 > 15 °C during the night (for review, see Farkas and Zajácz, 2007). Figures 3A and 4B  
394 show a drop in median mass gains for linden and combined linden/chestnut flow in  
395 years 2014 – 2016. Average colony density did not change much: quick calculation for  
396 linden flow shows average density of 68 colonies/1000 m<sup>3</sup> for years 2011, 2012 and  
397 2013 against 74 colonies/1000m<sup>3</sup> in years 2014, 2015, 2016. Most likely alternative  
398 could be weather conditions, e.g. severe storms. Unfortunately, we have no  
399 precipitation or other weather-related data for most of the locations.

400

401 Attempts to limit the colony numbers at the popular locations are being fiercely  
402 objected by local beekeeping community as pointless during main flow. Our study  
403 demonstrates that resources, in our case growing stock, define the reasonable colony  
404 density at the location. In related study, environment carrying capacity was investigated  
405 in Saudi Arabia, an arid country where vegetation outside oases is not particularly rich.  
406 Two most popular flow sources are *Acacia tortilis* and *Ziziphus spina-christi*. The  
407 authors calculated availability of the resources by dividing number of flowering plants  
408 with number of the colonies, concluding that increase of the colony numbers in the last  
409 20 years more than halved the harvest in some cases (Al-Ghamdi et al., 2016).

410

411 A neighboring country, Croatia, has reported 11,500 tons of honey harvested on  
412 roughly 56,600 km<sup>2</sup> from 406,000 colonies (7.2 colonies/km<sup>2</sup>). Calculations returned

413 28.3 kg/colony and 2.03 kg/ha (European Commission, 2015). The hectare yield is  
414 more than double with roughly 10% less colonies per square kilometer. The question is  
415 whether increasing colony density would increase honey harvest per hectare of surface  
416 or is the carrying capacity already reached? In terms of geography, climate and  
417 vegetation, Slovenia is perhaps more similar to another neighbor, Austria. Official  
418 statistics of EU shows that Austria reported 370,000 colonies - 4.4 colonies/km<sup>2</sup> – and  
419 4,300 tons of harvested honey in 2015, giving 11.6 kg/colony or 0.51 kg/ha (European  
420 Commission, 2015). This is roughly half the density and half the yield per hectare in  
421 comparison with Slovenia, which, as noted above, reported similar honey yield per  
422 colony (12.8 kg). In Austria at half of the density the average colony yield is already at  
423 the level of Slovenia. Therefore, we speculate that domestic increase of colony number  
424 to increase honey yield is probably not the option. On the other hand, decreasing the  
425 colony numbers in Slovenia might improve the yield of individual beekeeper. Out of  
426 10,000+ registered beekeepers there are less than hundred professionals, making living  
427 exclusively out of honey production, which gives additional weight to such argument.

428

429 Majority of forest owners value their woodland property mostly as the source of timber.  
430 In hypothetical case, there are 244 colonies within 3 km radius at 8.64 colonies/km<sup>2</sup> in  
431 Slovenia. In case of acacia honey harvest of 10 kg per hive and back-yard retail price  
432 of 10 €/kg, this represents revenue of 24,400 €. On the other hand, black locust timber  
433 is valued at 50 €/m<sup>3</sup> (Slovenian State Forests pricelist, April 2019). Assuming  
434 consistent growing stock of 282 m<sup>3</sup>/ha within 3 km radius, maximum revenue of one  
435 hectare of black locust forest would reach one-time value of 14,100 €. After such  
436 consideration, forest owners might be encouraged to become stakeholders in  
437 beekeeping operations.

438

### 439 Competition for resources

440

441 Last but not least, honey bees share available resources with other species.  
442 Overcrowding environment with honey bee colonies could have consequences on wild  
443 bees (bumblebees and other bees not belonging to genus *Apis*) and non-bee pollinators.  
444 The researchers seem to be divided on the topic: one study, for example, found no direct  
445 competition between honey bees and wild bees for forage, e.g. wild bees mostly depend  
446 on the coverage of non-cultivated vegetation in Central Europe (Steffan-Dewenter &  
447 Tschardtke, 2000). On the other hand, Thomson (2004) showed honey bees (*Apis*  
448 *mellifera*), non-native to the New world, as a threat to native pollinators from genus  
449 *Bombus*. Another report made by Roubik and Wolda (2001) show that in native bees  
450 on the island in the Panama Canal did not suffer from introduction of Africanized *Apis*  
451 *mellifera*. Yet none of the above studies considered such high colony numbers at the  
452 location.

453

### 454 Conclusions

455

456 Woodlands and forests are important resources for honey bees and beekeepers.  
457 Quantity of the resource – e.g. growing stock around the apiary – is the most important  
458 parameter when considering the carrying capacity of the environment in terms of honey  
459 bee colony numbers. Consequently, colony density requires both ecological and  
460 economic considerations. Both have a common denominator: (too high) density has  
461 negative consequences for both beekeepers and most likely for other nectar-feeders as  
462 well. While relationships between nectar flows and colony mass gain is clear, the

463 relationships between honeydew flows and colony mass seem to be more complex and  
464 not as clear. Most likely, another variable should be built in the equation, explaining  
465 the yield: population of dew-producing insects. Nevertheless, our models could be used  
466 to develop recommendations for management of beehive density, providing that there  
467 are both colony density and growing stock data available.

468

## 469 Acknowledgments

470

471 We are indebted to Aleš Bozovičar and Jure Justinek from Slovenian Beekeepers'  
472 Association for help with locations and discussions about measured data. We are also  
473 grateful to Jože Glad who assisted with maps and to Jernej Bubnič, Ajda Moškrič and  
474 Barbara Piškur for critical reading. Last but not least, thanks for creative discussions go  
475 to Mitja Piškur from Slovenian State Forests.

476

477 This work was supported by the Slovenian Research Agency Programs P4-0133,  
478 Slovenian Research Agency grant V4-1807 and Breeding program for carniolan honey  
479 bee "*Apis mellifera carnica*".

480

## 481 Declaration of interest

482

483 Declarations of interest: None.

484

485 Literature

486

- 487 1. Al-Ghamdi A, Adgaba N, Getachew A, Tadesse Y. 2016. New approach for  
488 determination of an optimum honeybee colony's carrying capacity based on  
489 productivity and nectar secretion potential of bee forage species. *Saudi J Bio*  
490 *Sci*, 23(1): 92 – 100. DOI: 10.1016/j.sjbs.2014.09.020
- 491 2. Bitterlich W. 1948. Die Winkelzahlprobe. *AUg.forst- u. holzw. Ztg* 59, 4-5.
- 492 3. Brodschneider R, Crailsheim K. 2010. Nutrition and health in honey bees.  
493 *Apidologie*, 41: 278 – 294. DOI: 10.1051/apido/2010012
- 494 4. Brus, R. 2007. Medovite rastline – pravi kostanj. *Slovenski čebelar*, 5: 163-164.
- 495 5. Brus, R. 2012. Drevesne vrste na Slovenskem. 2. dopolnjena izd. Ljubljana:  
496 samozal., 406 pp.
- 497 6. Chikamai B and Tchatat M and Tieguhong J and Ndoye O. 2009. Forest  
498 Management for Non-Wood Forest Products and Services in Sub-Saharan  
499 Africa. *Discovery and Innovation*, 21: 3. DOI: 10.4314/dai.v21i3.48213.
- 500 7. Crane et al., 1987 Crane, E., Walker, P., Day, R. (1984): Directory of important  
501 world honey sources. London, International Bee Research Association: 384 str.
- 502 8. Crane E, Walker P. 1985. Important Honeydew Sources and their Honeys. *Bee*  
503 *World*, 66(3): 105 – 112.
- 504 9. Dornhaus A., Klügl F., Oechslein C., Puppe F., Chittka L. 2006. Benefits of  
505 recruitment in honey bees: effects of ecology and colony size in an individual-  
506 based model. *Behav Ecol*, 17(3): 336 – 344. DOI: 10.1093/beheco/arj036
- 507 10. ESRI 2018. ArcGIS Desktop: Release 10.6. Redlands, CA: Environmental  
508 Systems Research Institute.
- 509 11. European Commission. 2015. Honey market presentation.



- 510 12. Farkas A, Zajácz E. 2007. Nectar production for the Hungarian honey industry.  
511 European J Plant Sci Biotech 1(2): 125 – 151.
- 512 13. Frisch K von. 1965. Tanzsprache und Orientierung der Bienen. Springer, Berlin  
513 Heidelberg New York.
- 514 14. Galton D. 1971. Survey of a thousand years of beekeeping in Russia. Bee  
515 Research Association, London, England.
- 516 15. Gary NE, Witherell PC, Marston JM. 1978. Distribution and foraging activities  
517 of honey bees during almond pollination. J Apicul Res, 17(4): 188 – 194. DOI:  
518 10.1080/00218839.1978.11099926
- 519 16. Jašmak K. 1973. Medonosno bilje. pp. 237. Publisher: Kosta Jašmak, Beograd.
- 520 17. Kohl PL, Rutschmann B. 2018. The neglected bee trees: European beech forests  
521 as a home for feral honey bee colonies. Peer J, 6: e4602. DOI:  
522 10.7717/peerj.4602
- 523 18. McLellan AR. Honeybee colony weight as an index of honey production and  
524 nectar flow: a critical evaluation. J App Ecology 14(2): 401 – 408. DOI:  
525 10.2307/2402553
- 526 19. Meikle WG, Rector BG, Mercadier G, Holst N. 2008. Within-day variation in  
527 continuous hive weight data as a measurement of honey bee colony activity.  
528 Apidologie 39: 694 – 707. DOI: 10.1051/apido:2008055
- 529 20. Nicolson SW, Human H. 2008. Bees get a head start on honey production. Biol  
530 Lett. 4: 299 – 301. DOI: 10.1098/rsbl.2008.0034
- 531 21. Official Gazette of the Republic of Slovenia. 2009. Pravilnik za slovenski med  
532 z zaščiteno geografsko označbo. OGRS, 46, p 6324.

- 533 22. Oleksa A, Gawroński R, Tofilski A. 2013. Rural avenues as a refuge for feral  
534 honey bee population. *Journal of Insect Conservation* 17(3):465–472. DOI:  
535 10.1007/s10841-012-9528-6
- 536 23. Persano Oddo L, Piro R. 2004. Main European unifloral honeys: descriptive  
537 sheets. *Apidologie* 35: S38 – S81. DOI: 10.1051/apido:2004045
- 538 24. Pirnat, J. 2017. *Krajinska ekologija: univerzitetni učbenik*. Ljubljana:  
539 Biotehniška fakulteta, Oddelek za gozdarstvo in obnovljive gozdne vire, 206  
540 pp.
- 541 25. Redei K, Meilby H. 2000. Effect of thinning on the diameter increment in black  
542 locust (*Robinia pseudoacacia* L.) stands. *Silva Gandavensis*, 65.
- 543 26. Rihar J. 1992. *Mana iglavcev. Napovedovanje gozdnega medenja*. Pansan,  
544 Ljubljana.
- 545 27. Roubik DW, Wolda H. 2001. Do competing honey bees matter? Dynamics and  
546 abundance of native bees before and after honey bee invasion. *Population*  
547 *Ecology*, 41(1): 53 – 62. DOI: 10.1007/pl00012016
- 548 28. Seeley TD. 2007. Honey bees of the Arnot Forest: a population of feral colonies  
549 persisting with *Varroa destructor* in the northeastern United States. *Apidologie*  
550 38(1):19–29. DOI: 10.1051/apido:2006055
- 551 29. Slovenian Forest Service. 2015. National forest inventory data, databases  
552 ODSSESDV.dbf.
- 553 30. Statistical Office RS. 2018. Beekeeping. Yearly honey production.  
554 <https://www.stat.si/statweb>.
- 555 31. Steffan-Dewenter I., Kuhn A. 2003. Honeybee foraging in differentially  
556 structured landscapes. *Proc. R. Soc. London B*, 270: 569 – 575. DOI:

- 557 32. Steffan-Dewenter I. and Tscharntke T. 2000. Resource overlap and possible  
558 competition between honey bees and wild bees in central Europe.  
559 *Oecologia*, 122(2): 288—296. DOI: 10.1098/rspb.2002.2292
- 560 33. Thomson, D. 2004. Competitive interactions between the invasive  
561 European honey bee and native bumble bees. *Ecology*, 85(2): 458 – 470.  
562 DOI: 10.1890/02-0626
- 563 34. Visscher PK, Seeley TD. 1982. Foraging Strategy of Honeybee Colonies in a  
564 Temperate Deciduous Forest. *Ecology*, 63(6): 1790-1801.
- 565 35. Waddington, K. D. Visscher, K. P. Herbert, T. J. Richter, M. R. 1994.  
566 Comparisons of forager distributions from matched honey bee colonies in  
567 suburban environments. *Behavioral ecology and sociobiology*, 35: 423-429.  
568 DOI: 10.1007/bf00165845
- 569
- 570