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1 **Evaluation of 213 nm laser as an affordable alternative for the green elemental**  
2 **characterization of particulate matter on quartz fibre filters by laser ablation**  
3 **ICPMS**

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23 and providing data on PM and elemental concentrations.

24        **Abstract**

25        Airborne particles, denoted as particulate matter (PM), are one of major environmental pollutants.  
26        Particles smaller than 10  $\mu\text{m}$  ( $\text{PM}_{10}$ ) penetrate into the human lungs during breathing and exert damage  
27        by physical and chemical mechanisms. PM mass concentration in the air as well as major toxicants  
28        contained are thus regulated by relevant directives all around the world and their continuous monitoring  
29        is prescribed. Elemental composition of PM is one of the most often measured air-quality parameters  
30        and the standard method for its determination produces huge amounts of toxic chemical waste.  
31        Sustainable alternatives are thus sought for, such as laser ablation inductively coupled plasma mass  
32        spectroscopy (LA-ICPMS), which allows for the direct analysis of PM collected on a filter without the  
33        microwave-assisted extraction step prior to the analysis with ICPMS. In this work we evaluated the  
34        performance of 213 nm Nd:YAG laser system (LA213) compared to the more powerful 193 nm excimer  
35        laser (LA193) for this application, in order to facilitate the replacement of the standard method (i.e.  
36        MW/ICPMS) with a new, waste-free one. We show that LA213 produces good results when operated  
37        under optimized instrument conditions, which were in fact very similar to the LA193 system. Sensitivity  
38        for some elements was, however, a bit poorer, but this can be overcome with additional fine-tuning, if  
39        necessary. Wrapping-up our thorough evaluation we can conclude that the more affordable LA213 is  
40        suitable for air-quality monitoring purposes.

41

42        **Keywords:**

43        PM, trace elements, metal concentration, sustainable, air quality monitoring, LA-ICPMS

## 44 **1 Introduction**

45 Particulate matter (PM) is one of the major air pollutants threatening the environment and human  
46 population (World Health Organization, WHO). PM is responsible for millions of deaths every year,  
47 among others causing pulmonary and cardiovascular diseases, and cancer (Lelieveld et al., 2015; Roy et  
48 al., 2016; Zhang et al., 2022). On the other hand, PM contributes to climate change by interfering with  
49 sun and Earth radiation pathways (Fuzzi et al., 2015).

50 The excess of PM which disrupts the natural balance is emitted by anthropogenic activities, such as  
51 traffic, industry, and biomass burning (Chatoutsidou et al., 2019; Karagulian et al., 2015; Trippetta et al.,  
52 2016; Tsiouri et al., 2015). Although, due to its nanoparticle nature, PM is harmful on its own, different  
53 sources produce distinctive particle structures and compositions, which finally determine the level of PM  
54 toxicity (Goix et al., 2016; Hu et al., 2012). Typically, PM is composed of inorganic ions, organic carbon,  
55 elemental carbon, crustal elements, and trace metals, some of these being known toxicants, such as  
56 polycyclic aromatic hydrocarbons (PAH) and particular trace elements. Besides PAH, airborne As, Cd, Ni,  
57 and Pb are regulated by the EU Commission (Directive 2004/107/EC and 2008/50/EC) prescribing their  
58 limit values and assuring safe living conditions along with limiting their influence on the environment.  
59 Moreover, reduction of national emissions of some other trace elements has also been commanded to  
60 date (Directive (EU) 2016/2284).

61 Accordingly, a standard method (microwave digestion (MW) pre-treatment followed by inductively  
62 coupled plasma mass spectroscopy (ICPMS) or graphite furnace atomic absorption spectrometry (GF-  
63 AAS) analysis) for the element determination in PM is prescribed (EN14902:2005). It is certified for the  
64 regulated four elements, although monitoring of Cr, Cu, Se, and Zn has also been ordered in Directive  
65 (EU) 2016/2284 due to their adverse effects to human health and the environment (Jaishankar et al.,  
66 2014; Tchounwou et al., 2012). The main disadvantage of the standard method is the sample preparation

67 step. Microwave assisted digestion with concentrated nitric acid and hydrogen peroxide does not result  
68 in complete digestion of quartz fibre filter and several authors report that accurate multielement  
69 determination in various PM matrixes is not guaranteed by this method (Aldabe et al., 2013; Swami et  
70 al., 2001). Moreover, MW is an additional step to the analysis, which increases the likelihood of sample  
71 contamination and/or loss of volatile elements. It is time consuming and requires high-purity reagents to  
72 assure sufficiently low limits of detection (LOD). The required reagents not only raise the cost of analysis  
73 but also end up as a toxic waste, which is not negligible if considering every-day monitoring at numerous  
74 stations around the world (Bai et al., 2022; Gerboles and Buzica, 2008; Ogrizek et al., 2022; Pant et al.,  
75 2019; Rovelli et al., 2018; Wang et al., 1998).

76 A greener or at least waste-free solution to this problem is to use analytical techniques capable of  
77 determining elemental composition directly on the solid PM samples (Gini et al., 2021). Relevant  
78 techniques (*i.e.* X-ray fluorescence (XRF) and its variants, instrumental neutron activation analysis (INAA),  
79 particle induced X-ray emission (PIXE), and LA-ICPMS) which all have their advantages and drawbacks are  
80 described in detail and compared in our recent review paper (Ogrizek et al., 2022), in which energy  
81 dispersive XRF (EDXRF) and LA-ICPMS have been recognized as most promising to replace the standard  
82 method currently in use. LA-ICPMS generally provides lower LODs than EDXRF and thus allows to detect  
83 and quantify more elements especially in low-polluted samples, which is often necessary when  
84 performing source apportionment studies. Moreover, LA-ICPMS can be directly applied to quartz fibre  
85 filters, the most universal and widely used filter matrix including in environment agencies, since quartz  
86 withstands high temperatures needed for complementary organic and elemental carbon analyses.

87 While the standard MW/ICPMS and the proposed LA-ICPMS method can utilise identical ICPMS  
88 instruments, allowing monitoring agencies to keep exploiting the original ICPMS function for liquid  
89 environmental samples (*e.g.* ground water samples), save on space and eliminate/limit the need for

90 additional educated personnel, a new LA system is indeed a bigger initial investment compared to a  
91 microwave oven. The recently developed method for the elemental determination of PM on quartz fibre  
92 filters by Ogrizek et al. (2021) uses a 193 nm excimer LA system (LA193); however, more affordable  
93 alternatives also exist on the market, as is 213 nm Nd:YAG (LA213). While method limitations pertaining  
94 to the sample itself (e.g. PM inhomogeneity) have already been addressed in our previous paper (Ogrizek  
95 et al., 2021) and do not change with the laser wavelength, different LA systems and wavelengths do not  
96 couple with material in the same way, which needs to be explored for the case of PM samples.

97 While excimer LA193 systems are characterized by homogeneous energy distribution inside the laser  
98 beam generating evenly flat craters, solid state LA213 systems are known for irregularities in laser  
99 energies inside the laser beam, resulting in a more undulating crater base (Gilbert et al., 2014). For this  
100 reason, in analytics, LA193 systems are usually preferred. Besides, less fractionation effects are observed  
101 in the case of high energy lasers compared to the higher wavelength lasers, which improves analytical  
102 accuracy especially when homogeneous, matrix-matched standards are not available for calibration.

103 Another effect arising from the different wavelengths used for laser ablation is distinct size distribution  
104 of particles generated during the ablation, which are generally moved towards the smaller size range (<  
105 150 nm) in the case of LA193 compared to the LA wavelengths above 200 nm (particle sizes of up to 1  
106  $\mu\text{m}$  are anticipated; both sizes are related to the ablation of NIST 610 series silicate glasses) (Guillong et  
107 al., 2003). Larger particle sizes can be problematic, since they tend to deposit close to the ablation crater  
108 and may not get transported to the ICPMS at all. Furthermore, even if these particles enter the plasma,  
109 their conversion to ions may be incomplete, which consequently influences the sensitivity of the  
110 measurement. It has been further shown that sample characteristics, *i.e.* sample transparency also  
111 influence particle size distribution and stability of the signal (Gonzalez et al., 2002; Guillong and Gunther,  
112 2002). While LA193 performed better in the case of more transparent samples (*e.g.* NIST 612 and 614),

113 for more opaque NIST 610 the particle size distribution and signal stability were similar for LA193 and  
114 LA213 systems (Guillong et al., 2003).

115 As PM samples are opaque, beige to dark brown in colour, no differences in fractionation effects due to  
116 ablation are expected between LA193 and LA213. Nevertheless, another fractionation effect could be  
117 due to different laser irradiances or power densities, which can cause sample heating or melting  
118 (Wohlgemuth-Ueberwasser and Jochumb, 2015). The 193 nm excimer and 213 nm Nd:YAG are both  
119 nanosecond lasers, where melting was observed in the past compared to the femtosecond lasers (Luo et  
120 al., 2020; Wohlgemuth-Ueberwasser and Jochumb, 2015). If similar pulse width (4–5 ns) and laser energy  
121 density were used, LA213 produced more melting during the ablation of sulphides compared to LA193  
122 (Wohlgemuth-Ueberwasser and Jochumb, 2015). Notwithstanding, with different pulse width and  
123 fluence applied in the ablation of alloy steels (3 ns with fluences 3.5–12.1 J cm<sup>-2</sup> for LA213 and 15 ns with  
124 fluences 5.1–14.5 J cm<sup>-2</sup> for LA193), the least melting was observed in the case of short pulse width and  
125 low fluence with LA213 (Luo et al., 2020). Note, however, that with LA213 systems higher fluence is  
126 usually needed for comparable ablation to the LA193 (*e.g.* 4.2 vs 2.7 J cm<sup>-2</sup> for sulphide analysis (Gilbert  
127 et al., 2014), or 15 vs 5 J cm<sup>-2</sup> for typical reference materials (RM) as are NIST 610, GSE-1G, MACS-3, and  
128 STDP5 (Jochum et al., 2014), for LA213 vs LA193, respectively).

129 In this study we investigate utilisation of the 213 nm Nd:YAG LA-ICPMS for the determination of the  
130 elemental composition of PM collected on quartz fibre filters. A new method is developed for this  
131 particular application and compared to the already existing 193 nm excimer LA-ICPMS (Ogrizek et al.,  
132 2021). The potential of the more affordable LA213 system for air-quality monitoring purposes is  
133 specifically addressed, which would facilitate the worldwide transition to a more sustainable analysis in  
134 the future. Sensitivity of both methods in respect to different elements is explored and applicability of  
135 the new method is discussed, considering also the initial investment in the equipment, which can be

136 critical especially in research, while for environment agencies the return of investment is faster due to  
137 circumventing the laborious pre-extraction step (>3 h per sample set) and huge amounts of toxic waste  
138 (100 ml per sample).

## 139 **2 Experimental section**

### 140 **2.1 Samples and standard materials**

141 Ambient PM<sub>10</sub> samples were donated by the Slovenian Environment Agency (ARSO) and were collected  
142 during their routine monitoring according to the current EU legislation (EN12341:2014 and  
143 EN14902:2005). Sampling was performed by a high volume sampler (Digitel DHA-80 Aerosol Sampler,  
144 Switzerland) for 24 h with a flow rate of 500 L min<sup>-1</sup> on quartz fibre filters (PALL, USA; filter diameter 150  
145 mm). The samples were collected in Maribor, Slovenia (111,000 inhabitants), representative of an urban  
146 environment with traffic and biomass burning as two major pollution sources. Samples from various  
147 seasons thus provided different filter loadings and analyte concentrations.

148 Trace elements in glass SRM 610 and 612 (NIST, USA) were used for testing the intraday stability of the  
149 instruments and comparison of sensitivity between LA systems. Pristine quartz fibre filters (PALL, USA)  
150 were used as filter blanks.

### 151 **2.2 LA-ICPMS instruments**

152 LA is a special process used to introduce solid samples to ICPMS for analysis. Since no sample pre-  
153 treatment is required, PM<sub>10</sub> samples were simply cut and mounted to a microscopic glass with a double  
154 sided tape (Scotch, USA) and put into a holder for analysis. Two LA systems using different laser  
155 generators and wavelengths were compared: (i) 193 nm nanosecond excimer LA system (193 nm ArF\*;  
156 Analyte G2, Teledyne Photon Machines Inc., Bozeman, MT; later referred as LA193) at the National  
157 Institute of Chemistry, Slovenia, and (ii) 213 nm nanosecond solid state LA system (Nd:YAG, LSX 213 G2;

158 Teledyne Photon Machines Inc., Bozeman, MT; later referred as LA213) at the Masaryk University, Czech  
159 Republic. 2-volume ablation cell with He as a carrier gas was used in both cases (HelEx II ablation cell for  
160 LA193 and HelEx for LA213). Carrier gas flow rate was 0.5 and 0.6 L min<sup>-1</sup> for the cup, respectively, and  
161 0.3 L min<sup>-1</sup> for the cell (in both cases). The same type of a quadrupole ICPMS instrument (Agilent 7900,  
162 Agilent Technologies, Santa Clara, CA) was used for detection in both cases, with Ar as a makeup gas (0.8  
163 L min<sup>-1</sup>). Kinetic energy discrimination (KED) process included in the instrument was used for successful  
164 minimization of polyatomic interferences.

165 Twenty elements (Al, As, Ba, Cd, Co, Cr, Cu, Fe, K, Mn, Mo, Ni, Pb, Rb, Sb, Sn, Ti, Tl, V, and Zn) were  
166 measured simultaneously with the total acquisition time of 0.5 s and element specific dwell times as  
167 listed in Table S1.

### 168 **2.2.1 Parameter optimization**

169 Although it is known that each airborne particle is unique, the focus of the current paper is on the  
170 analysis of bulk concentrations of trace elements in PM deposited on collection filters, as required by the  
171 legislation. Therefore, instrument parameters were optimized to measure large sample area in a short  
172 amount of time, and to ablate the majority of the PM deposited on the filter rather than obtaining a  
173 perfect LA image. Optimization of LA193 parameters is described in Ogrizek et al. (2021) and LA213  
174 optimization was approached in a similar way.

175 Based on our previous knowledge, ablation was performed using the adjacent line-scan mode, with the  
176 largest beam size LA system is capable to produce, and scan speed calculated to provide square pixels  
177 with a minimal smear. Acquisition time was synchronized with the washout time of the used ablation cell  
178 and was similar in both LA systems.

179 Three parameters were optimized experimentally: fluence (0.9–8.6 J cm<sup>-2</sup>), scan speed and repetition  
180 rate that correspond to dosage 2–20 were used with the LA213 (details in Table S2). During fluence

181 optimization, two ablation passes were repeated over the same ablation line to determine the amount  
182 of residual PM on a filter. Due to limitations of the maximum repetition rate (20 Hz), dosage optimization  
183 had to be performed at two different scan speeds ( $250 \mu\text{m s}^{-1}$  for dosages 2–10 and  $125 \mu\text{m s}^{-1}$  for  
184 dosages 12–20); therefore, a sum of pixels was used to compare the different dosages instead of an  
185 average as used in the fluence optimization. When no signal increase was observed at increasing  
186 dosages, we assumed that the relevant PM mass had been ablated from the filter. An area of  $3.15 \text{ mm}^2$   
187 was ablated for every tested parameter and experiments were repeated on four different PM samples  
188 (two summer and two winter samples) for all twenty elements.

### 189 **2.2.2 Sample Analysis**

190 The set of optimized LA213 parameters was first applied to the ablation of a blank filter, to study the  
191 potential influence of impurities to the PM composition determination.

192 Eight ambient  $\text{PM}_{10}$  samples with different PM loadings and elemental concentrations (Table S3) were  
193 further measured by both LA systems using the optimal sets of parameters to compare instrument  
194 performance. Due to the inhomogeneity of PM material deposited on a filter, every time a randomly  
195 selected  $3 \times 12 \text{ mm}^2$  filter area was ablated, which has been previously shown representative of the  
196 whole PM filter (Ogrizek et al., 2021). Statistical evaluation of the smallest representative filter area for  
197 LA193 is described in detail in the respective paper by Ogrizek et al. (2021) and results in 30 minutes of  
198 analysis time per sample.

199 The samples were quantified based on known elemental concentrations from the standard MW/ICPMS  
200 analysis (data received by the Slovenian Environment Agency) with a leave-one-out cross-validation  
201 procedure (LOOCV). When heteroscedastic characteristics of the data drastically influenced the  
202 determination of low concentration samples, a weighted ( $x^{-1}$ ) linear model was used (*i.e.* for As, Ba, Cd,

203 Co, Cr, Cu, Fe, Mn, Ni, Pb, Rb, V, Zn). Reasons for why we used such an approach for calibration are given  
204 in our method development paper (Ogrizek et al., 2021).

205 Differences between LA systems were statistically evaluated with the paired t-test or the paired  
206 Wilcoxon test in the case of elements where the assumption of normality was not met (the latter was  
207 assessed with the Shapiro-Wilk test). Furthermore, all three determination methods (LA193, LA213,  
208 MW/ICPMS) were compared with the repeated measures ANOVA or its appropriate non-parametric  
209 alternative (*i.e.* Friedman test). For visual representation of uncalibrated results the data were  
210 normalized (1) to allow for the comparison of different scales.

$$x_{normalized} = (x - average) stdev^{-1} \quad (1)$$

211 Statistical analysis and data handling was performed in software R using raster, data.table, reshape2  
212 (Wickham, 2007), tidyverse (Wickham et al., 2019), caret, and rstatix packages.

213 The sensitivity of both LA systems was compared by visual assessment of the capability of distinguishing  
214 between the ablation signal and the features of gas blank (*i.e.* background noise of the instrument), and  
215 graphical comparison of the measured counts. Additionally, NIST glass standards 610 and 612 were used  
216 for this purpose (ablation parameters are listed in Table S4). Note, signal intensities (counts) of NIST  
217 glasses are more prone to the ablation of different depths as PM samples with the majority of PM  
218 material captured in the upper layers of the filter. The ratio between the two LA systems was calculated  
219 for every element in order to i) compare ablation volumes and ii) better evaluate method sensitivity.

## 220 3 Results and discussion

### 221 3.1 LA213 parameters optimization

222 The LA213 system was first optimized to assure the finest possible sensitivity and repeatability. As  
223 already mentioned in the introduction part, how the laser interacts with a sample depends on sample  
224 characteristics (*e.g.* transparency, density) and laser wavelength. Usually, LA213 needs more energy  
225 (higher fluence) compared to LA193 for similar ablation. For example, NIST 600 glass series RM is  
226 optimally ablated using fluence 3–5 J cm<sup>-2</sup> with LA193, whereas 8–12 J cm<sup>-2</sup> is needed with LA213. Our  
227 samples, however, are much more complex due to their two-component structure (filter matrix and  
228 deposited PM), which does not allow to easily draw parallels with RM or other samples from the  
229 literature; instrument parameters thus have to be carefully optimized for the specific application.

230 In PM on a filter samples, particles are mostly deposited in top layers of the filter and are not strongly  
231 bound to filter fibres. As PM represents only a small proportion of the sample mass (sample = filter +  
232 deposited PM), impurities in the filter matrix strongly interfere with the analysis of trace PM  
233 concentrations if large amount of filter matrix is ablated. Since we are only interested in the composition  
234 of PM and not in the filter matrix itself, the optimal fluence will be strong enough to allow for the  
235 ablation of the deposited PM but low enough to ablate as little quartz fibres as possible at the same  
236 time. Therefore, relatively low fluence is preferred for the ablation of PM samples, with the correction to  
237 higher dosages (*i.e.* more overlapping shots) in case that PM still remains on the filter after the selected  
238 number of overlapping shots.

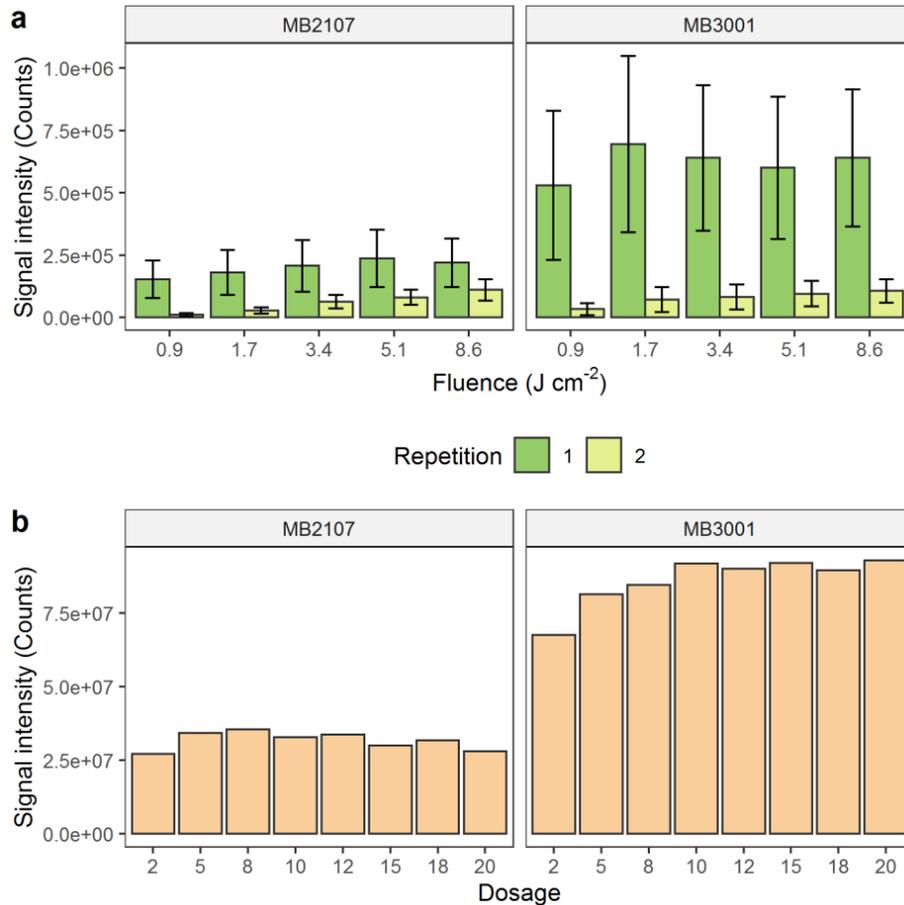
239 **Table 1** Optimized instrument parameters for the ablation of PM deposited on quartz fibre filters

Parameter	LA193*	LA213
<b>Pre-determined parameters by educated guess</b>		
Laser beam size	128 µm (square mask)	125 µm (square mask)

Scan speed	256 $\mu\text{m s}^{-1}$	250 $\mu\text{m s}^{-1}$
Acquisition time	0.5 s	0.5 s
<b>Optimized parameters</b>		
Dosage	10	10
Repetition rate	20 Hz	20 Hz
Fluence	1.5 $\text{J cm}^{-2}$ (with 10x demagnification)	1.7 $\text{J cm}^{-2}$
Ablated area	3 × 12 $\text{mm}^2$ (random sampling)	3 × 12 $\text{mm}^2$ (random sampling)

240 \*from Ogrizek et al. (2021)

241 It turned out that optimal instrument parameters are very similar for both LA systems (Table 1). In the  
242 case of LA213, 1.7  $\text{J cm}^{-2}$  was chosen as the optimal fluence, because higher fluences did not provide  
243 significant increase in the signal intensities (Fig. 1a). The observed increase in the signal intensity with  
244 increasing fluence especially in the second ablation pass can be attributed to the ablation of quartz fibre  
245 material together with its impurities at higher fluences, which should be avoided. This explanation is  
246 additionally confirmed with the lack of signal intensity increase at higher dosages (Fig. 1b). Therefore, we  
247 selected dosage 10 as sufficient for winter samples, whereas for summer samples even less dosage  
248 would be adequate.



249

250 **Fig. 1** Optimization of LA213 parameters: example for Cr. One highly loaded sample (MB3001) and one sample with a low PM

251 mass-load (MB2107) are shown. In the optimization of fluence, dosage was fixed at 10, and for the optimization of dosage,

252 fluence 1.7 J cm<sup>-2</sup> was used. Error bars in A represent variability between the pixels (standard deviation). In B, the sum of pixels

253 is shown instead of an average, not allowing us to calculate the standard deviation

254 We should mention at this point that instrument parameters are optimized for typical ambient PM

255 concentration levels in Europe (PM<sub>10</sub> mass concentration < 50 µg m<sup>-3</sup>). However, if samples are collected

256 in much more polluted areas with significantly higher mass loadings, the dosage needs to be adjusted,

257 keeping the fluence at the determined level (*vide supra*). From this perspective, LA213 has some

258 limitations, since the maximum repetition rate is 20 Hz, compared to LA193 where 300 Hz or even 1000

259 Hz can be used (Šála et al., 2020). For instance, with the maximum repetition rate of 20 Hz, scan speed

260 has to be adjusted for dosages above 10, resulting in a prolonged measurement time.

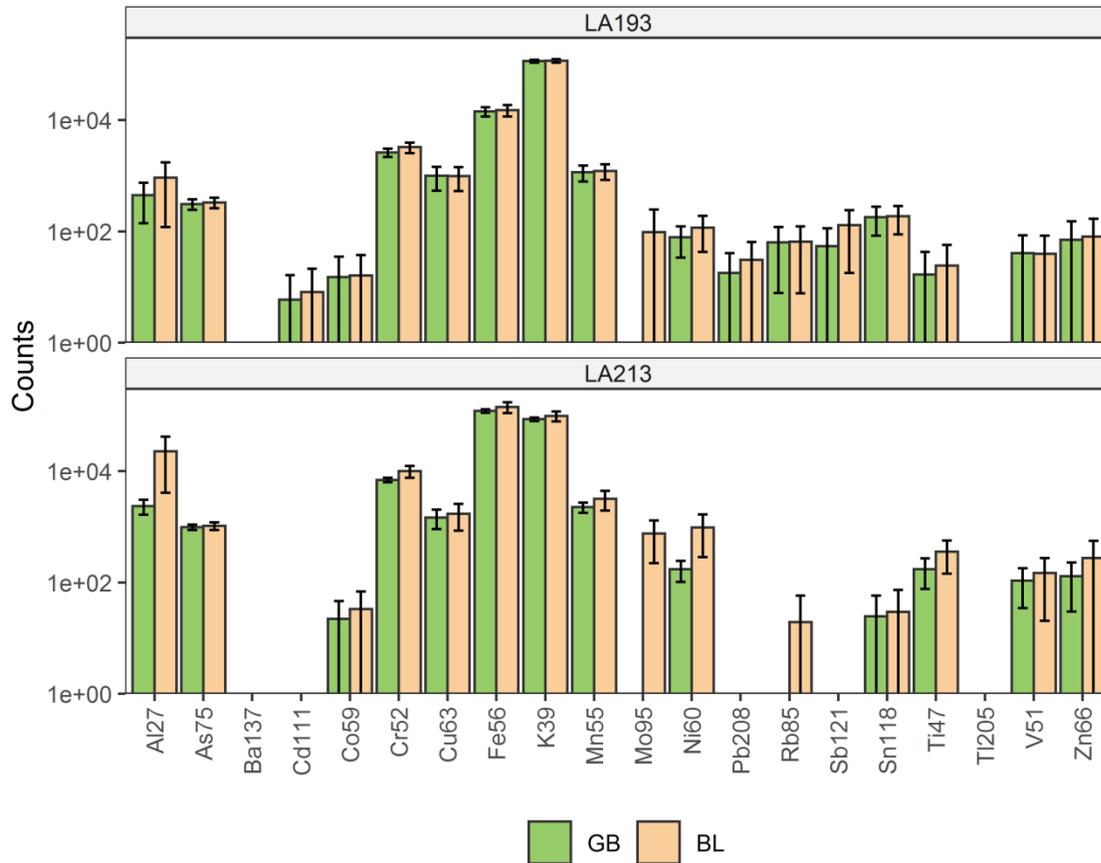
### 261 3.2 Blank filter ablation

262 Although ablation parameters were optimized to ablate as little filter matrix as possible, in the upper  
263 layers of the filter quartz fibres and PM are intertwined together so that ablation of some fibres cannot  
264 be completely avoided. Along with the fibres, some impurities from the filter are always sampled as well.  
265 Whenever the impurities of blank filter differ significantly from the background noise (*i.e.* gas blank),  
266 their contribution needs to be accounted for during PM analysis. Therefore, filter blanks were analysed  
267 with both LA systems using the set of optimized instrument parameters and compared (Fig. 2).

268 The most pronounced interferences originating from blank filters that were detected with both LA  
269 systems are Al, Cr, and Mo (Fig. S1). Aluminium impurities are not problematic, since Al is one of the  
270 Earth crust elements that are naturally present in PM in relatively high concentrations even in low-  
271 polluted areas. On the other hand, Cr and Mo are less abundant in PM and in low-polluted samples non-  
272 negligible blank signals can substantially affect their determination.

273 Overall, with use of the optimised set of parameters, LA213 seems to ablate more blank filter material  
274 and consequently more impurities than LA193 (observe the differences between gas blanks and filter  
275 blanks in Fig. S1). With LA213, Ni was clearly detected in blank filters and some barely noticeable  
276 differences between gas and filter blanks were also found for Fe, K, Mn, Ti, and Zn. On the other hand,  
277 LA193 barely detected Ni, Sb, and Ba impurities, whereas the latter two elements were not detected  
278 with LA213 at all, indicating higher detection limits.

279 Among the listed elements, special attention should be paid to Cr, Ni, and Mo when using LA213 (filter  
280 blanks accounted for up to 20, 20, and 45% of low-concentration samples, respectfully), and to Cr when  
281 using LA193 (filter blank accounted for 8% of the low-concentration sample). All other impurities were  
282 low enough compared to sample concentrations to avoid an influence on the final quantification  
283 (impurities in blanks contributed <3% to the lowest sample signal in both LA systems).



284

285 **Fig. 2** Gas blank (GB) vs. filter blank (BL): Comparison of signals from the background noise (*i.e.* GB) with the impurities in blank  
 286 filters (*i.e.* BL). The plotted error bars are standard deviations between the pixels, representing their scattering

287 **3.3 PM sample ablation: comparison of LA213 with LA193**

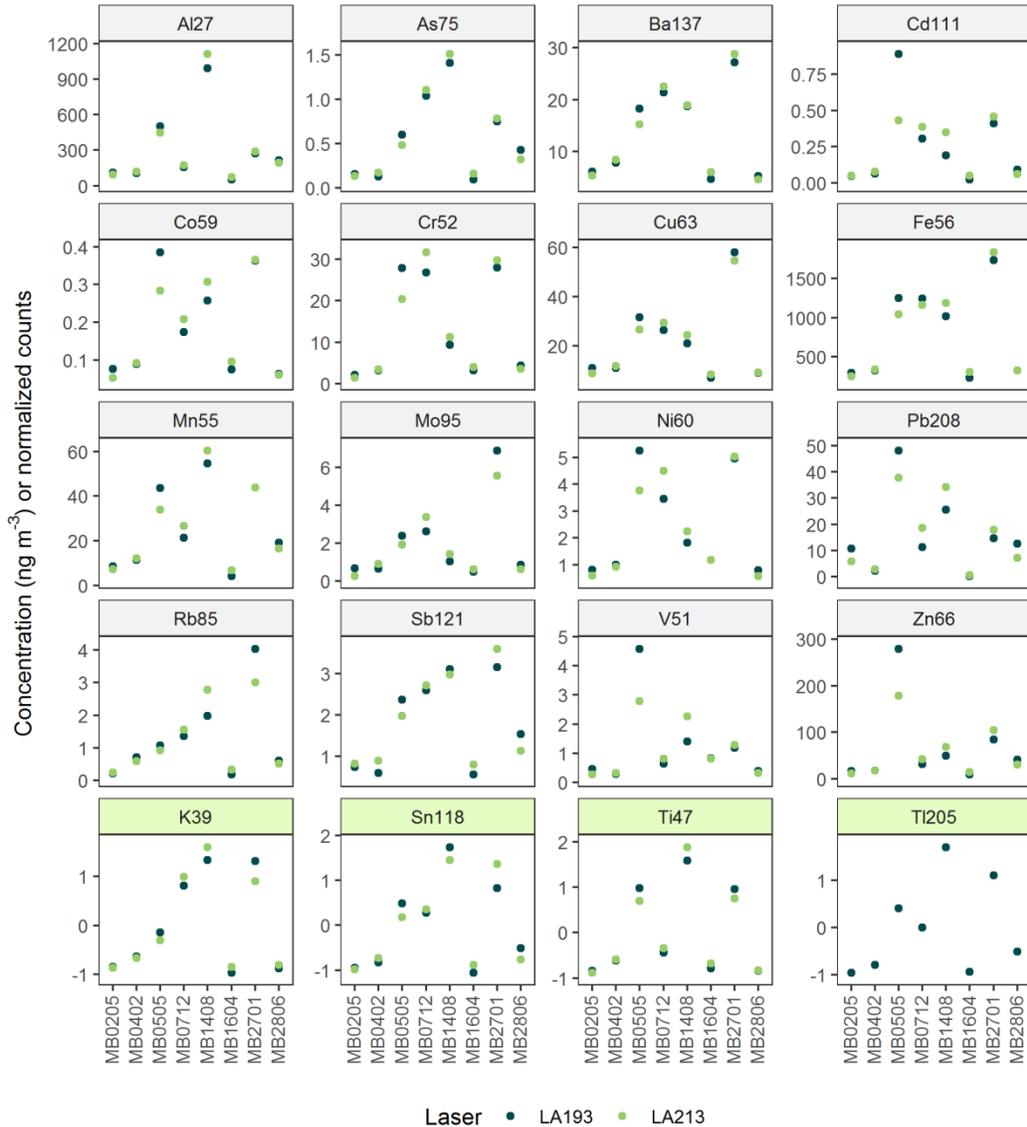
288 It has already been explained that laser interaction with the ablated material is wavelength specific and  
 289 can lead to distinctive elemental fractionation effects. The ablation of PM samples is performed in a line-  
 290 scan mode and since the majority of PM mass is deposited in the upper layers of the sample, an in-depth  
 291 fractionation effect is not likely to be observed. Moreover, potential fractionation effects are expected to  
 292 be corrected for by calibration with matrix-matched standards. However, calibration still remains an  
 293 issue in LA-ICPMS.

294 Since PM material on quartz fibre filter is a very specific sample in its structure, none of the commonly  
295 used RM (*e.g.* glass or gelatine standards, pressed pellets, spiked filter matrix, ...) is suitable for its  
296 calibration (Brown et al., 2013; Miliszkievicz et al., 2015; Šala et al., 2017). Matrix-matched RM of PM on  
297 quartz fibre filter with certified elemental composition is also not yet commercially available. Although  
298 the use of ambient PM samples previously characterized with MW/ICPMS as standards for the  
299 calibration is not an ideal practice, since erroneous determination due to the MW pre-treatment step  
300 can lead to inaccurate LA-ICPMS analysis, it is a matrix-matched material and fits the purpose of our  
301 research.

302 The constructed calibration curves were not ideal (note that MW/ICPMS used for their standardization is  
303 not certified for all elements); but still, LOOCV analysis showed they were comparable for LA193 and  
304 LA213 (Fig. S2). Both LA systems exhibited high  $R^2$ , low root mean squared errors (RMSE), and low mean  
305 absolute errors (MAE) for Al, As, Co, Cu, Fe, Mn, Mo, Sb (Fig. S2). Somehow poorer performance was  
306 obtained for Cd, Rb, and Zn, with both instruments. Furthermore, we can see that LOOCV calibration  
307 curves, which are presented in Fig. S3 for the case of As, show substantial deviation from the rest of the  
308 curves only when the highest concentration measured is omitted (as part of statistical procedure). This  
309 may lead to biases in the quantification of analytes with concentrations in the upper part of each  
310 calibration curve, while confidence of the determination of low-concentration samples should not be  
311 compromised.

312 In Fig. S4 we compare the results of sample analyses with all three methods. Samples determined with  
313 both LA systems and MW/ICPMS do not reveal any credible evidence for the difference for any of the  
314 elements, which was also confirmed by statistical testing (Table S5). Note that MW/ICPMS results were  
315 used as calibration standards, masking eventual systemic biases due to incomplete digestion in the  
316 respective method. Airborne concentrations of all elements as determined by LA193 and LA213 are

317 further compared in Fig. 3, Table S6, and Table S7. Higher deviations (although randomly distributed)  
318 between the LA systems are observed in the samples with the highest element concentrations, which is  
319 in line with the nature of the constructed calibration curves as explained in the previous paragraph.  
320 Moreover, the highest concentrations measured fell outside of the calibration range, which resulted in  
321 even poorer confidence of their determination. Consequently, high absolute differences between the  
322 two LA systems are observed at the highest concentrations measured. Note that according to Directive  
323 2004/107/EC, accuracy of determination of As, Cd, and Ni in PM is reported within 40 % by MW/ICPMS.  
324 From all this we may conclude that the performance of LA213 is comparable to LA193.



325

326 **Fig. 3** LA193 vs LA213: Comparison of ambient sample element concentrations. The top 4 rows (grey) present calibrated

327 elements with concentration on y-axis, and the bottom row (green) are non-calibrated elements with normalized counts on y-

328 axis, since the 4 elements (*i.e.* K, Sn, Ti, and Tl) were not measured with MW/ICPMS. Tl could only be detected with LA193

329 Although we showed that both LA systems perform comparably well in terms of accuracy, different

330 sensitivity was observed for some elements. LA213 turned out to be less sensitive for Ba, Cd, Pb, Sb, Sn

331 and Tl in PM samples (Fig. S5) as well as NIST glass standards (Fig. S6). Sb and Sn are not considered

332 problematic, since they are usually present in PM in fairly high concentrations. On the contrast, in low-

333 concentration samples, Ba, Cd, and Pb were at the edge of LOD in LA213 measurements, which was not  
334 the case with LA193 (refer here also to Fig. S7). We were also unable to determine TI with LA213 in  
335 either the PM samples or NIST glass standards. Among the referenced elements in NIST 610, TI has the  
336 lowest reported concentration (approximately 7-times lower than the other elements of interest), which  
337 points to the poor sensitivity of the method for this particular element. Besides, TI is also volatile and  
338 could be lost during the ablation, especially with LA213. Note, however, that when particular elements  
339 are of specific interest, high LODs and/or low sensitivities can be overcome by additional optimization  
340 and introducing more material for their analysis (*e.g.* bigger beam size or longer acquisition times; the  
341 latter consequently means measuring fewer elements at a time) (van Elteren et al., 2019).

342 Although we observed a similar pattern in sensitivity towards different elements for real samples and  
343 NIST standards, we should not neglect the fact that with the different lasers we indeed ablated different  
344 volumes of the material due to different beam sizes (256  $\mu\text{m}$  for LA193 and 125  $\mu\text{m}$  for LA213), but also  
345 due to differences in the used fluence and laser wavelengths, influencing intensity of the signal (*i.e.*  
346 counts). As already mentioned, ablation volume is especially important in the case of NIST standards.  
347 Therefore, in order to understand differences between both ablations, we further examined the ratios  
348 between the elements measured with the LA systems in both RM, which are shown in Table S8.

349 Based on the ratios of most elements (*i.e.* Al, As, Co, Cr, Cu, Fe, K, Mn, Ni, Ti, V, and Zn), larger volumes  
350 seem to be ablated with LA213 compared to LA193. If this is true, LA213 indeed exhibits much lower  
351 sensitivity for the above-mentioned elements. Furthermore, when comparing the ratios for NIST 610 and  
352 612, more abnormalities were found in the case of NIST 612 (see Table S8), which implies more  
353 fractionation effects in the case of NIST 612. Similar observations were reported by Guillong et al. (2003),  
354 where fractionation indexes of LA193 and LA213 were different in the case of NIST 612, but still very  
355 similar in the case of NIST 610.

356 Ba, Cd, Pb, Sb, Sn, and Tl are normally produced by anthropogenic activities (*e.g.* traffic, industry) and  
357 usually exhibit lower airborne concentrations than elements originating from natural sources (*e.g.* Al, Fe,  
358 K). In source apportionment studies, those 6 elements are usually associated with traffic or traffic-  
359 related emission factors as are brake and tyre wear (*e.g.* Ba, Cd, Pb, Sb, Sn), biomass burning (*e.g.* Cd, Tl),  
360 or industrial emission factors as are metallurgical emissions (*e.g.* Pb, Tl) (Almeida et al., 2020; Belis et al.,  
361 2013; Ferreira et al., 2016; Landis et al., 2017; Massimi et al., 2020). Our samples were collected in an  
362 urban environment, close to a road with heavy traffic where anthropogenic emissions are relatively  
363 constant regardless of the seasons. Nevertheless, with LA213 we could barely detect some of these  
364 elements in summer samples, which may be generally problematic when studying samples collected in  
365 warmer parts of the year or in rural areas with less traffic and industrial emission sources. From this  
366 perspective, LA193 is a preferential choice of use in source apportionment studies conducted in low-  
367 polluted areas or where Tl is the targeted element of interest (*e.g.* determination of the influence of  
368 specific industry on air pollution), unless LA213 method parameters are further optimized for these  
369 elements (Karbowska, 2016).

370 Even though the sensitivity of LA213 is poorer than of LA193 for some elements, it is still sufficient to  
371 meet the current legislation (limit values are 6, 5, 20, and 500 ng m<sup>-3</sup> for As, Cd, Ni, and Pb, respectively;  
372 EU Directive 2004/107/EC and 2008/50/EC), and to compete with the current standard method (low  
373 limit of working range is set at 0.5, 0.1, 2, and 1 ng m<sup>-3</sup> for As, Cd, Ni, and Pb, respectively). In our work,  
374 airborne concentrations were in the range of 0.09–1.5 ng m<sup>-3</sup> for As, 0.02–0.9 ng m<sup>-3</sup> for Cd, 0.6–5.3 ng  
375 m<sup>-3</sup> for Ni, and 0.3–50 ng m<sup>-3</sup> for Pb as shown in Fig. 3; we were able to determine them all, although  
376 some concentrations fall below the certified LOD of MW/ICPMS. Multi-element determination and other  
377 pros of the method can be exploited as comparable with MW/ICPMS, without concerning about the  
378 problems connected with MW digestion in the case of LA-ICPMS. Therefore, we can conclude that LA213

379 performs suitably well for monitoring of elemental concentrations in PM according to the current  
380 legislation and is a valid alternative to the more expensive LA193 system.

#### 381 **4 Conclusions and outlook**

382 Striving for sustainability has become an important part of our everyday life. Analytical techniques often  
383 produce large quantity of toxic chemical waste and to achieve sustainable future, green analytical  
384 methods have to be promoted even in research laboratories. Air quality monitoring is one of the  
385 continuous processes where green analytical methods are especially appreciated. The intention of  
386 detecting pollution for the purpose of its mitigation and at the same time producing toxic waste and  
387 polluting the environment seems somehow contradictory. Moreover, monitoring network is expected to  
388 continuously grow and extend even to the less developed countries, which will multiply the number of  
389 analyses and consequently their effect on the environment. Waste-free analytical methods with a  
390 minimal burden to the environment should thus replace the old methods, whenever possible.

391 One such example is the proposed LA-ICPMS for the elemental characterization of PM instead of the  
392 standardized MW/ICPMS method. While LA-ICPMS certainly circumvents the production of toxic waste  
393 (approx. 140 ml 16 % nitric acid pH <1) and minimizes manual work (by 4-times), one may question the  
394 consumption of gases and power due to the longer operation time during the analysis (about 10-times  
395 longer compared to solution ICPMS). Although a detailed Life cycle analysis is out of scope of this  
396 manuscript, we still want to argue in favour of the method we promote also from this point of view. By  
397 comparing the cost of consumables, which somehow resembles their production costs including energy  
398 consumption, LA-ICPMS turns out 2-times cheaper than MW/ICPMS. If we further account for MW  
399 digestion, the difference in power consumption between LA-ICPMS and MW/ICPMS is only left at 2-3  
400 times and is still expected to decrease, because modern LA instruments tend to be more eco-friendly.  
401 While the development of LA systems is currently on the rise, no drastic improvements are expected in

402 the field of ICPMS, although N<sub>2</sub> plasmas are on the horizon and will positively affect both MW/ICPMS and  
403 LA-ICPMS analyses. Moreover, there is already 20 % difference in power consumption between the two  
404 instruments used, in favour of the LA213, and ICPMS usually uses more energy when working in solution  
405 mode compared with LA (Fietzke and Frische, 2016). Considering also necessary requirements for the  
406 specific application (*i.e.* large laser beams, moderate laser energy), we believe that power consumption  
407 of LA systems can be still substantially improved.

408 Many different LA systems are currently available on the market, varying in performance and price  
409 availability. We show that the 193 nm excimer LA system enables low-limit determination of all 20  
410 elements of interest and is presented as a preferable choice in source apportionment studies, where  
411 multi-element quantification is especially appreciated. On the other hand, 213 nm solid state LA system  
412 exhibits slightly higher LODs for some elements, which can be critical in low-polluted samples.  
413 Nevertheless, LA213 is still shown satisfactory for use in the elemental determination of PM according to  
414 the current legislation (*i.e.* meets prescribed limit values) and it can be further fine-tuned for specific  
415 target analytes or extremely low-concentration samples while retaining the benefits over other surface  
416 techniques (*e.g.* low LOD, direct measurement on fibrous filters). Therefore, we assess the cheaper  
417 LA213 system adequate for air-quality monitoring purposes, which lowers the initial investment needed  
418 for the transition to the greener analytical alternative.

419 In this paper we present state-of-the-art capabilities of LA-ICPMS systems for the analysis of ambient PM  
420 samples and identify their potential benefits in air quality monitoring. However, just as standard  
421 MW/ICPMS has many drawbacks (*e.g.* incomplete digestion, limited RM), LA-ICPMS also has some  
422 limitations. We want to encourage relevant manufacturers to direct their development to the production  
423 of standards/parts specific for this application enabling proper method validation and better  
424 automatization. The field of LA-ICPMS is continuously improving and actively developing in the direction

425 of having more and more commercially available certified matrix matched standards for various  
426 calibration purposes, and autosamplers customized for specific samples. This is necessary for assuring  
427 the required traceability in regulatory agencies and facilitating the analysis of large number of samples.  
428 We believe LA-ICPMS is the green analytical method on the rise, that can contribute with a small step  
429 towards achieving the desired sustainable future even within air-quality monitoring.

430

## 431 **Declarations**

432 **Ethics approval and consent to participate** Not applicable. No human or animal subject were involved in  
433 the study.

434 **Consent for publication** Not applicable.

435 **Availability of data and material** The datasets generated during and/or analysed during the current  
436 study are available from corresponding author on reasonable request.

437 **Competing interests** The authors have no relevant financial or non-financial interests to disclose.

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447

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